

Phytoremediation of Metals in the Environment for Sustainable Development

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Phytoremediation is an environmental cleanup strategy in which selected green plants are employed to remove, contain, or render environmentally toxic contaminants harmless. This is an emerging biogeotechnological application based on the "green liver concept" and operates on the principles of biogeochemical cycling. Phytoremediation projects have been successfully implemented in the developed nations for the cleanup of metal polluted /contaminated soil/water including the restoration of degraded mines. Extensive diversity of native and non-native plants has been used in this strategy. This article would focus the relevance of phytoremediation technology for cleanup of metals in soils for sustainable development including some recent trends. Bioconcentration, phytoextraction, phytomining, hyper accumulation; natural accumulators, enhanced accumulation by using selected group of plants and soil amending chelators; phytostabilization, phytotransformation, and rhizofiltration; Rhizosphere biotechnology of metal hyper accumulating naturally occurring and transgenic plants and role of mycorrhizal fungi (both ecto- and arbuscular mycorrhizae) are emphasized in this review. Phytoremediation is being used along with or, in some cases as substitute of expensive conventional chemical and mechanical cleanup methods. The *pros* and *cons* of phytoremediation technology are also discussed.

Key Words: Biodiversity prospecting, Metal hyperaccumulators, Metal transporters, Metal tolerance, Phytoextraction, Phytoremediation, Phytostabilization, Transgenics, Phytovolatilization, Rhizofiltration

Introduction

Even the supposedly most pristine environment in Arctic circle and Antarctica has not been spared by the globally transported anthropogenic pollutants/contaminants of three major groups viz, i) toxic trace metals (includes metalloids and radio nuclides) ii) acidifying gases (SO_x), and iii) variety of persistent organic pollutants that play a major role in global climate change (AMAP 2002). This presentation would concentrate only on "trace metals, metalloids and radionuclides" which would intercept biogeochemical cycles (figure 1); hence, is a subject of utmost concern to "Man and Biosphere". Air (Atmosphere), water (Hydrosphere) and soil (Lithosphere) are not only the reservoirs of nutrients but also act as receptacles for a variety of inorganic and organic pollutants (figure 1). Noosphere is a man-dominated ecosystem, which releases variety of inorganic and organic pollutants. Phytosphere (green belts) is the sink for these pollutants.

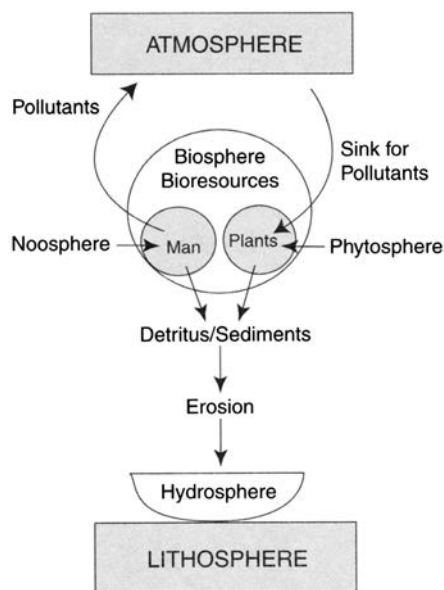


Figure 1 Concept of biogeochemical cycle on which bioremediation relies

Phytoremediation technology is based on use of plants to cleanup metals; metalloids and radionuclide contaminated/polluted sites both in developed and transition nations (Bell et al. 1988, Iskandar 2002, Iskandar & Kirkham 2001, Kabata-Pendias 2001). This, strategy can be broadly termed as "Bioremediation" which involves soil micro flora and fauna. Bioremediation is being commercialized owing to its success in many real world ecosystems (Comis 1995, Cunningham & Berti 1993a,b). It is emerging as an innovative tool encompassing Biology, Geology, Biotechnology and allied disciplines with cutting edge applications for sustainable development (Cunningham & Ow 1996, Cunningham et al. 1995, Vangronsveld & Cunningham 1998, Groudev et al. 2001a,b, Mejare & Bulow 2000, Prasad & Freitas 1999, 2003). This review evaluates the global findings, primarily dealing with higher plants, focusing the salient developments, and linking the related branches of contemporary biotechnology highlighting the new levels of understanding. Suggestions for possible further research are also mentioned.

Though the phytodiversity and metal polluted sites are enormous in India, there have not been many research investigations from India in the area of metal tolerance in higher plants, biodiversity prospecting for mineral exploration and cleanup of metals in the environment (Aery & Tiagi 1986, Aravind & Prasad 2003a,b, Bhati & Singh 2003, Chandra Sekhar & Puvvada 1997, Chandra Sekhar et al. 1998, 2001, 2002, 2003a,b, Dhillon & Dhillon 2003, Prasad, 1995, 1996, 1997, 1998, 1999, 2001, 2002a,b, 2003, 2004a,b, Prasad et al. 2001a,b, Rama Kumar & Prasad 1999, Reddy & Prasad 1989, 1990, 1993, 1995). Contrastingly countries with less plant diversity have generated enormous data on metal detoxification mechanisms and their application for phytoremediation by appropriate field and laboratory experiments and currently this knowledge is being commercialized (Baker et al. 1995, Brooks 1998, Chaney et al. 1997, Cunningham & Ow 1996, Cunningham et al. 1995, Dushenkov et al. 1995, 1997a,b, 1999, Ebbs & Kochian 1998, Entry et al. 1999, Ernst 2000, Glass 1999, 2000, Greger & Landberg 1999, Hossner et al. 1998, Huang & Cunningham 1996, Kramer & Chardonnnes 2001, Kumar et al. 1995, Ma et al. 2001a,b, McGrath & Zhao 2003, Palmer et al. 2001, Pilon-Smits & Pilon 2000, 2002, Prasad & Freitas 1999, 2000a,b, 2003, Pulford & Watson 2003, Raskin & Ensley 2000, Rugh et al. 1996, Ruiz & Romero 2002, Salt et al.

1995a-c, 1998, Schnoor 2000, Vangronsveld & Cunningham 1998, Wong 2003).

The following are some of the technical terms connected with phytoremediation ("phyto" meaning plant, and the Latin suffix "remedium" meaning to clean or restore) of metals in the environment (figure 2):

1. *Phytoremediation*: Use of plants to remediate contaminated soil, water and air.
2. *Phytoaccumulation*: The uptake and concentration of contaminants (metals or organics) within the roots or aboveground portion of plants.
3. *Phytoaccumulation coefficient*: Metal concentration in plant dry matter/extractable metal in soil.
4. *Phytoextraction*: The use of plants at waste sites to accumulate metals into the harvestable, above ground portion of the plant and, thus, to decontaminate soils.
5. *Phytoextraction coefficient*: The ratio of extractable metal concentration in the plant tissues (g metal/g dry weight tissue) to the soil concentration of the metal (g metal/g dry weight soil).
6. *Phytomining*: Use of plants to extract inorganic substances from mine ore.
7. *Phytostabilisation*: Plants tolerant to the element in question are used to reduce the mobility of elements, thus, they are stabilised in the substrate or roots.
8. *Phytovolatilization*: The uptake and transpiration of a contaminant by a plant, with release of the contaminant or a modified form of the contaminant to the atmosphere from the plant.

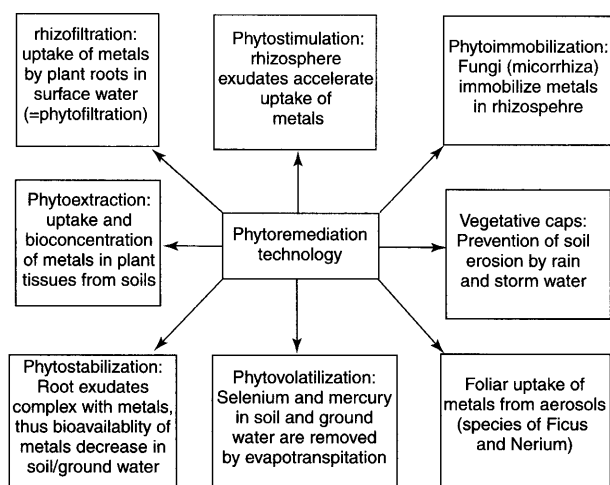


Figure 2 Various processes of phytoremediation technology

9. *Rhizofiltration/Phytofiltration*: Roots or whole plants of element accumulating plants absorb the element from polluted effluents and are later harvested to diminish the metals in the effluents
10. *Rhizosecretion*: A subset of molecular farming, designed to produce and secrete valuable natural products and recombinant proteins from roots

Phytoremediation involves diverse use of plants for *in situ* treatment of metal contaminated soils, sediments, water and air. It is best applied at sites with shallow contamination of metals (Bishop 1997, Bizily et al. 1999, Boyajian & Carreira 1997, Brennan & Shelley 1999, Brooks et al. 1998, Brown 1995, Brown et al. 1994, Chen et al. 2000, 2003, Conte & Piccolo 2003, Ebbs & Kochian 1997, Ebbs et al. 1997, Entry et al. 1996, Fitz & Wenzel 2002, Flathman & Lanza 1998, Francesconi et al. 2002, Garbisu & Alkorta 2001, Gisbert et al. 2003, Glass 1999, 2000, Gleba et al. 1999, Greger & Landberg 1999, Greger & Wang 2003, Hamlin 2002, Heaton et al. 1998, Hossner et al. 1998, Huang 1997, Huang et al. 1997, 1998, Kamnev & van der Lelie 2000, Karenlampi et al. 2000, Lanza & Flathman 2001, Lasat 2002, Lasat et al. 1997, 1998a,b, 2002, Linger et al. 2002, Madrid et al. 2003, McGrath & Zhao 2003, McGrath et al. 2001, McIntyre & Lewis 1997, Meagher 2000, Mejáre & Bülow 2000, Nedelkoska & Doran 2000a,b, Nie et al. 2002, Ouyang 2002, Palmer et al. 2001, Prasad 2003c, 2004a,b, Prasad & Freitas 1999, 2000a,b, 2003, Rangsayatorn et al. 2002, Raskin et al. 1994, 1997, Robinson et al. 1997a,b, Römkens et al. 2002). A model representing phytoremediation of metals by inducing metal tolerance or avoidance in plants (phytostimulation/phyto-transformation) over a period of time is depicted in figure 3.

International Conference on the Biogeochemistry of Trace Elements (ICOBTE) is one of the important international events dedicated to this area, discusses the interdisciplinary science in this field, linking phenomena observed in the biogeosphere to physical and chemical reactions in the pedo- and lithosphere. So far, 7 international conferences have been held as follows: Orlando, USA (1990), Taipei, Taiwan (1993), Paris, France (1995), Berkeley, USA (1997), Vienna, Austria (1999) Guelph, Canada (2001) and Uppsala, Sweden (2003). Battelle, an USA based environmental consultancy agency organizes another important series of conference on the theme *in situ* and onsite remediation technology covering inorganics and organics.

Metal Accumulating Plants

Hyperaccumulator plants are geographically distributed and are found throughout the plant kingdom (Brooks 1998, Chaney et al. 1995). To date approximately 400 taxa, ranging in growth habit from annual herbs to perennials are known. Hyper accumulator plants have been identified on all continents, both in temperate and tropical environments (table 1). Natural occurrences of hyperaccumulators for Ni include New Caledonia, Cuba, Southeast Asia, Brazil, southern Europe and Asia Minor; for Zn and Pb include northwest Europe; and for Cu and Co include south-central Africa. Some families and genera are particularly well documented for Ni [Brassicaceae (*Alyssum* & *Thlaspi*), Euphorbiaceae (*Phyllanthus*, *Leucocroton*) and Asteraceae, Zn Brassicaceae (*Thlaspi*), and Cu and Co (Lamiaceae, Scrophulariaceae) (Brooks 1998). There are not many Cr hyper accumulators in nature, but there are numerous Ni hyper accumulators. A few Cr hyperaccumulators have been identified, partly because Cr exists predominantly in the +3 oxidation state and is very insoluble and much less available for plant uptake.

The ability of a plant to hyperaccumulate any one metal may infer some ability to accumulate other metals (Prasad & Freitas 2003). Some metals may interact competitively for accumulation (e.g., Zn and Ni in calamine and serpentine soils). The number of Ni hyper accumulator taxa are more than 300 in about 35 families (table 2). They commonly have 3-4% Ni in dry matter of leaves. *Alyssum betolonii*, which is endemic to serpentine soils, is known for its high concentration of

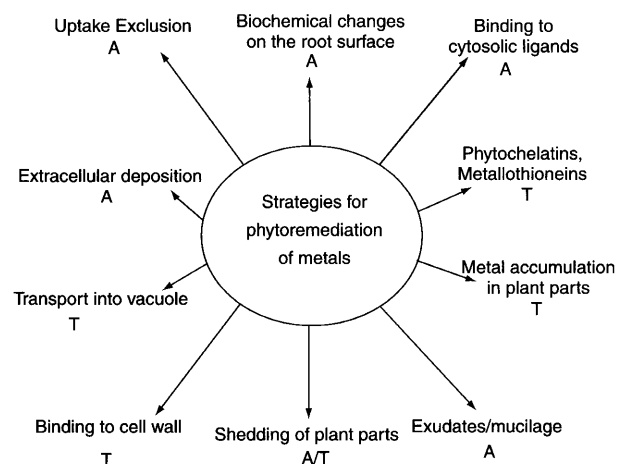


Figure 3 Avoidance and tolerance strategies (ecophysiological basis for the phytoremediation technology) exhibited by plants exposed to elevated concentrations of metals.

Table 1. Plants that accumulate Lead (Pb), Nickel (Ni) and (Zn) >1 mg g⁻¹ d.wt from Brassicaceae. Zinc Tabulation is based on ascending order of metal budget in plant metal wise (for details refer Palmer et al. 2001)

Metal	Taxa	Distribution	Reference
Lead (Pb)			
2740	<i>Thlaspiceraulescens</i> J.Presel & C.Presl	United Kingdom	Shimwell and Laurie 1972
8200	<i>Thlaspi rotundifolium</i> sub sp. <i>cepaefolium</i> (Wulfen) Rouy & Fouc	Central Europe	Reeves and Brooks 1983
Nickel (Ni)			
1050	<i>Cardamine resedifolia</i> L.	Italy	Vergnano Gambi and Gabbrielli 1979
1280	<i>Alyssum singarense</i> Boiss. & Hausskn.	Iraq	Brooks et al. 1979
2000	<i>Thlaspi bulbosum</i> Spruner ex Boiss.	Greece	Reeves and Brooks 1983
2440	<i>T.japonicum</i> H.Boissieu	Japan	Reeves 1988
3000	<i>T.epirotum</i> Halacsy	Greece	Reeves and Brooks 1983
3140	<i>Pseudosempervivum sempervivum</i> Boiss & Balansa) Pobed	Turkey	Reeves 1988
3420	<i>Alyssum tenium</i> Halacsy	Greece	Brooks and Radford 1978
3960	<i>A.fallacinum</i> Hausskn	Crete	Brooks and Radford 1978
4000	<i>Thlaspi ochroleucum</i> Boiss. & Heldr.	Greece	Reeves and Brooks 1983
4480	<i>Alyssum alpestre</i> L	Southern Europe	Brooks and Radford 1978
4550	<i>A.euboicum</i> Halacsy	Greece	Brooks and Radford 1978, Reeves et al. 1983a
4590	<i>A.obovatum</i> (C.A.Mey) Turez	Russia	Brooks et al. 1979
4900	<i>A.condensatum</i> Boiss. & Hausskn.	Iraq, Syria	Brooks et al. 1979, Reeves et al. 1983a
5530	<i>Thlaspi montanum</i> L. var. <i>montanum</i>	USA	Reeves et al.1983b
6230	<i>Alyssum virgatum</i> Nyar.	Turkey	Brooks et al. 1979
6600	<i>A.smolikanum</i> Nyar.	Greece	Brooks and Radford 1978
7080	<i>A.murale</i> Wealdst&Kit	Balkans	Brooks and Radford 1978 Brooks et al. 1979
7290	<i>A.oxycarpum</i> Boiss. & Balansa	Turkey	Brooks et al. 1979
7390	<i>A.giosnanum</i> Nyar.	Turkey	Brooks et al. 1979
7600	<i>A.peltarioides</i> subsp. <i>virgatiforme</i> Nyar.T.R.Dudley)	Turkey	Brooks et al. 1979, Reeves et al. 1983a
7700	<i>A.floribundum</i> Boiss.&Balansa	Turkey	Brooks et al. 1979
7860	<i>A.penwinensis</i> T.R.Dudley	Iraq	Brooks et al. 1979
52120	<i>Thlaspi cypricum</i> Bornm.	Cyprus	Reeves 1988
8170	<i>A.anatolicum</i> Nyar.	Turkey	Brooks et al. 1979
9090	<i>A. akamasicum</i> B.L.Burt	Cyprus	Brooks et al. 1979
10000	<i>A.serpyllifolium</i> Desf	Spain	Brooks et al. 1981,Dudley 1986b
10200	<i>A.bertolonii</i> subsp.scutarinum Nyar	Balkans	Reeves et al. 1983b
10200	<i>A.syriacum</i> Nyar.	Syria	Brooks et al. 1979
10400	<i>A.crenulatum</i> Boiss	Turkey	Brooks et al. 1979
10900	<i>A.callichroum</i> Boiss. & Balansa	Turkey	Brooks et al. 1979
11400	<i>Bornmulleria</i> sp petri Greuter, Charpipn & Dittrich	Greece	Reeves et al. 1983a
11500	<i>Alyssum eriophyllum</i> , Boiss. & Hausskn.	Turkey	Brooks et al. 1979
11700	<i>A.discolor</i> T.R.Dudley & Huber-Morath	Turkey	Brooks et al. 1979
11800	<i>Thlaspi tymphaeum</i> Hausskn.	Greece	Reeves and Brooks 1983
11900	<i>Alyssum trapeziforme</i> Nyar.	Turkey	Brooks et al. 1979
12000	<i>Thlaspi goesingense</i> Halacsy	Greece	Reeves and Brooks 1983
12400	<i>T.graecum</i> Jord	Greece	Reeves and Brooks 1983
12500	<i>Alyssum heldreichii</i> Hausskn.	Greece	Brooks and Radford 1978
12500	<i>A.robertianum</i> Bernard ex Godron & Gren	Corsica	Brooks et al. 1979
13400	<i>A.bertolonii</i> Desv.	Italy	Minguzzi and Vergnano 1948, Palmer et al. 2001
13500	<i>A.cilicium</i> Boiss. & Balansa	Turkey	Brooks et al. 1979, Reeves et al. 1983a
13500	<i>A. huber-morathii</i> T R Dudley	Turkey	Brooks et al. 1979
13600	<i>Thlaspi kovatsii</i> Heuffel	Yugoslavia	Reeves and Brooks 1983, Baker and Brooks 1989

(cont.)

Table 1 (contd.)

Metal	Taxa	Distribution	Reference
13700	<i>Alyssum markgrafii</i> O E Schulz	Albania	Brooks et al. 1979, Reeves et al. 1983a
14800	<i>Streptanthus polygaloides</i> A Gray	USA	Reeves et al. 1981
16200	<i>Thlaspi caerulescens</i> J Presl & C	Greece	Reeves and Brooks 198
16300	<i>Alyssum chondrogynum</i> B L Burt	Cyprus	Brooks et al. 1979
16500	<i>A.dubertretii</i> gomb	Turkey	Brooks et al. 1979
16500	<i>A.carcium</i> T.R.Dudley & Huber-Morath	Turkey	Brooks et al. 1979
17100	<i>A.troodii</i> Boiss.	Turkey	Brooks et al. 1979
17600	<i>Pseudosempervivum aucheri</i> (Boiss.) Pobed	Turkey	Reeves 1988
18100	<i>Alyssum constellatum</i> Boiss.	Turkey	Brooks et al. 1979
18300	<i>Thlaspi rotundifolium</i> (L.) Gaudin var. <i>corymbosum</i> (Gay)	Central Europe	Reeves and Brooks 1983a,b
18900	<i>Alyssum samariferum</i> Boiss. &Haussskn.	Samar	Brooks et al. 1979, Reeves et al. 1983a
18900	<i>Peltaria dumulosa</i> Post	Asia	Reeves et al. 1980
19200	<i>Bornmuellaria glabrescens</i> (Boiss.& Balansa) Cullen & T.R.Dudley	Turkey	Reeves et al. 1983a
19600	<i>Alyssum davisianum</i> T.R. Dudley	Turkey	Brooks et al. 1979
20000	<i>Alyssum cassium</i> Boiss	Turkey	Brooks et al. 1979
20800	<i>Thlaspi elegans</i> Boiss	Turkey	Reeves 1988
18300	<i>T. rotundifolium</i> (L.) Gaudin var. <i>corymbosum</i> (Gay)	Central Europe	Reeves and Brooks 1983
18900	<i>A.samariferum</i> Boiss. &Haussskn	Samar	Brooks et al. 1979, Reeves et al. 1983a
21100	<i>A.pinifolium</i> (Nyar.)T.R.Dudley	Turkey	Brooks et al. 1979, Reeves et al. 1983a
21300	<i>Bornmuellera baldaccii</i> (Degen) Heywood	Greece	Reeves et al. 1983a
22200	<i>Alyssum pterocarpum</i> T.R Dudley	Turkey	Brooks et al. 1979, Reeves et al. 1983a
22400	<i>A.lesbiacum</i> (P.candargy) Rech.f	Greece	Brooks et al. 1979
23600	<i>A.cypricum</i> Nyar	Cyprus	Brooks et al. 1979
24300	<i>A.masmenaeum</i> Boiss	Turkey	Brooks et al. 1979
26900	<i>Thlaspi jaubertii</i> Hedge	Turkey	Reeves 1988
27300	<i>T.caerulescens</i>	Germany, Belgium	Reeves and Brooks1983
29400	<i>Alyssum argenteum</i> All	Italy	Brooks et al. 1979, Reeves et al. 1983a
31000	<i>Thlaspi sylvium</i> (as <i>T.alpinum</i> subsp. <i>sylvium</i>)	Central Europe	Reeves and Brooks 1983
31200	<i>Bornmuellaria tymphaea</i> (Haussskn.) Haussskn	Greece	Reeves et al. 1983a
34400	<i>Peltaria emarginata</i> (Boiss.)Haussskn.	Greece	Reeves et al. 1980
35600	<i>Thlaspi oxyceras</i> (Boiss.) Hedge	Turkey, Syria	Reeves 1988
Zinc (Zn)			
1150	<i>Thlaspi idahoense</i> Payson	USA	Reeves et al. 1983b
1400	<i>Cochlearia pyrenaica</i> DC.	Belgium	Ramaut et al. 1972
1680	<i>Thlaspi violascens</i> Schott & Kotschy.	Turkey	Reeves 1988
3000	<i>Thlaspi montanum</i> L.	USA	Reeves et al. 1983b
3000	<i>T.ochroleucum</i> Boiss.&Helder.	Greece	Reeves and Brooks 1983
3090	<i>T.parvifolium</i> A.Nelson	USA	Reeves et al. 1983b, Reeves 1988
3520	<i>T. liaceum</i>	Turkey	Reeves 1988
3890	<i>T.magellanicum</i> Comm.Poir.	Argentina, Chile	Reeves 1988
10500	<i>Thlaspi bulbosum</i> Spruner ex Boiss.	Greece	Reeves and Brooks 1983
21000	<i>T.praecox</i> Wulfen	Bulgaria	Reeves and Brooks 1983
26700	<i>Arabidopsis thaliana</i> L	Pennsylvania,USA	Cannon 1960, Reeves 1988
16000	<i>Thlaspi stenocarpum</i> (boiss.) Hedge	Spain	Reeves and Brooks 1983
21000	<i>T.rotundifolium</i> subsp. <i>cepaefolium</i> (Wulfen)	Italy	Reeves and Brooks 1983
11000	<i>T.rotundifolium</i> (L.) Gaudin	France	Reeves and Brooks 1983
39600	<i>T. caerulescens</i> J Presl & C Presl	Germany	Linstow 1924,Reeves & Brooks 1983

Table 2. Plants that hyper accumulate metals and concentration limits (Brooks 1998, Baker et al. 1995, 1998, Hossner et al. 1998, Francesconi et al. 2002, Reeves & Baker 1998)

Metal	Concentration in leaves mg/g DW	No. of taxa	No. of families	Examples
Cd	> 0.1	1	1	<i>Thlaspi caerulescens</i>
Pb	> 1	14	6	<i>Minuartia verna</i>
Co	> 1	28	11	<i>Aeollanthus biformifolius</i>
Cu	> 1	37	15	<i>Aeollanthus biformifolius</i>
Ni	> 1	317	37	<i>Alyssum bertolonii</i> , <i>Berkheya coddii</i>
Mn	> 10	9	5	<i>Macadamia neurophylla</i>
Zn	> 10	11	5	<i>Sedum alfredii</i> <i>Thlaspi caerulescens</i>
As	> 22	2	1	<i>Pteris vittata</i> , <i>Pityrogramma calomelanos</i>

Ni (> 10,000 mg kg⁻¹ in leaves). The fact that serpentine (ultramafic) soils also contain other elements such as Cr has led to the assumption that the preferential accumulation of Ni in many species of *Alyssum* is due to a selective uptake mechanism (Sequeira & Pinto da Silva 1991).

Brassica juncea (Indian mustard) - a high-biomass producing plant can accumulate Pb, Cr(VI), Cd, Cu, Ni, Zn, Sr, B, and Se (Palmer et al. 2001, Prasad 2001, Kumar et al. 1995, Salt et al. 1998, Raskin & Ensley 2000). It produces biomass of over 20 times that of *Thlaspi caerulescens* (Salt et al. 1998). *Brassica juncea* had the best ability to transport lead to the shoots. *B. juncea* cultivars varied widely in their ability to accumulate Pb, with different cultivars ranging from 0.04% to 3.5% Pb accumulation in the shoots to 7 to 19% in the roots (Kumar et al. 1995). Except for sunflower (*Helianthus annuus*) and tobacco (*Nicotiana tabacum*), other non-Brassica plants had phytoextraction coefficients less than one.

Phytoremediation is being investigated and/or used commercially in a number of field scenarios (Glass 1999, 2000, Prasad 2001, 2003, 2004). Plant species are selected for phytoremediation based on their growth rates and yield, the depth of their root zone, and their ability to bioaccumulate contaminants and the type of rhizospheric changes they can initiate for decontamination (Prasad 2001, 2003, 2004). Variations in plant accumulation of metals are regulated by growth seasons, species and tissues of the plants (Benson et al. 1981, Ellis & Eslick 1997, Hagemeyer et al. 1994, Huang and Cunningham 1996). The ecophysiological, biochemical and molecular aspects underlying the success of phytoremediation are detailed in table 3 with key reference. Processes fostering phytoremediation are also enumerated in table 4 for specific cases of metals.

Phytoremediation of heavy metal contaminated soils in temperate regions has been reviewed recently by Pulford and Watson (2003). Tree crops suitable to tropical and sub-tropical ecosystems such as *Eucalyptus camaldulensis* were investigated for their important attributes such as tolerance to a wide range of soil type, and pH, rapid growth and high wood yield for timber and fuel wood. Mixed effluents were found to be beneficial for improving soil health and for enhancing growth and biomass production of the tree crops (Latimer et al. 1996, Eklund 1995, El-Hasan 2003, Hagemeyer et al. 1994, Landeen & Mitchell 1986, Prasad et al. 2001). This kind of practice may come up as an alternative forestry/agroforestry related water technology for irrigation in water and biomass scarce areas (Bhati & Singh 2003, Lepp 1975, Marcantonio et al. 1998, Pulford & Watson 2003, Schnoor 2000, Zeleznick & Skousen 1996).

Phytoextraction

In this process, a crop species known to accumulate metals, metalloids or radionuclides is cultivated on metalliferous soil. The crop is harvested and the contaminant is recovered (Bañuelos et al. 1999, Brooks et al. 1998, Conte & Piccole 2003, Ebbs & Kochian 1998, Garbisu & Alkorta 2001, Huang et al. 1997, Nedelkoska & Doran 2000a,b, Kumar et al. 1995, Prasad 1998, McGrath & Zhao 2003). In order to have a practicable treatment alternative, one needs a vigorously growing plant (>3 tons dry matter/ha/yr) that is easily harvested and which accumulates large concentrations of metal in the harvestable portion (>1000 mg/kg metal). As a general rule, readily bioavailable metals for plant uptake include cadmium, nickel, zinc, arsenic, selenium, and copper. Moderately bioavailable metals are cobalt, manganese, and iron; while lead, chromium, and uranium are not easily bioavailable.

Table 3. Ecophysiological, biochemical and molecular basis for phytoremediation

Mechanism	Key reference (s)
Mycorrhizae	Bi et al. 2003, Entry et al.1999, Jan Colpaert 2001 in Prasad 2001, Khan 2001, Khan et al.2000, Riesen and Brunner 1996, Shetty et al. 1994, 1995, El Kherbawy 1998
Cell wall exudates	Briat and Lebrun 1999, Hall 2002
Plasma membrane (passive uptake and active efflux)	Hall 2002, Clemens et al. 2002
Biodiversity prospecting	Cannon 1960, Brooks, 1998, Prasad and Freitas 2003 Ferritins, metallothioneins, Prasad 1995, 1997, 1999, Brait and Lebrun 1999, Rama Kumar and Prasad 1999
glutathione derived peptides (phytochelatins)	Hartley-Whitaker et al. 2001, Cobbet and Goldsbrough 2002, Rama Kumar and Prasad 1999, Rauser 1995, 2003, Mejare and Bulow 2000, Reddy and Prasad 1990, Schmoeger et al. 2000, Vatamaniuk et al.2000, Sanita di Toppi et al. 2002
Low molecular weight organic acids and amino acids	Hall 2002, Banks et al 1994, Brait and Lebrun 1999, Deiana et al. 2003, Delhaize et al. 1993, Hall 2002, Harmens et al. 1994, Huang et al. 1998, Kerkeb and Kramer 2003, Kramer et al.1996, Ma et al. 2001b, Larsen et al. 1998, Pigna et al. 2003, Wu et al.2003
Heat shock proteins	Reddy and Prasad 1993, Neumann et al. 1994, Reddy and Prasad 1995
Metal transporters (cation efflux family (formerly cation diffusion family)	Vacuolar compartmentation Clemens et al. 2002, Prasad 1995, 1997, 1999
Genetic and transgenic strategies for metal hyper accumulation	Assunção et al. 2001, Lasat et al.2000, Persans et al. 2001, Pence et al.2000, Guerinot 2000, Prasad 2002a, 2003a, Mc Grath and Zhao 2003
Tissue culture techniques	Clemens 2001, Clemens et al.2002, Karenlampi et al.2000, Maywald and Weigel 1997, Misra and Gedamu 1989, Pollard et al.2002, Raskin 1996, Rugh et al. 1996, Pena and Seguin 2001, Gisbert et al.2003, Krämer and Chardonnens 2001
Rhizosphere biotechnology and Physiology andbiochemistry of metal tolerance and toxicity	Brewer et al.1997, Nedelkoska and Doran 2000b, Shanks and Morgon 1999, Palmer et al. 2001
Over expression of glutathione (precursor for metal sequestration)	Wenzel et al. 1999 in Prasad and Hagemeyer 1999, Wenzel et al. 2003 in Prasad 2003c, Prasad and Strzalaka 2002
Limitations of chelate enhanced phytoextraction and concomitent limitations	Zhu et al. 1999a,b, Xiang et al. 2001 Blaylock et al. 1997, Huang 1997, Huang et al.1997, Kock and Mitchel 1957, McGrath and Zhao 2003 Karmner et al. 1996, Madrid et al. 2003, Prasad 2003c, Römkins et al. 2002, Wu 1999

Use of soil amendments such as synthetics (ammonium thiocyanate) and natural zeolites mobilize metal ions and increased the metal bioavailability to plants for the uptake and accumulation (Gersevich & Petunkin 1993, Zorpas et al. 1999). EDTA and NTA (nitrilotriacetic acid) etc. have been used as chelators for rapid mobility and uptake of metals from contaminated soils by plants. Thus, usage of synthetic chelators significantly increased Pb and Cd uptake and translocation from roots to shoots facilitating phytoextraction of the metals from low grade ores. On the other hand synthetic cross-linked polyacrylates, hydrogels have protected plant roots from heavy metal toxicity and prevented the entry of toxic metals into roots (Huttermann et al. 2003 in Prasad 2003c, Laperche et al. 1997).

Different genotypes of plants behave differently in the uptake of metals (Benson et al. 1981). Greger (2003) demonstrated this phenomenon using *Salix* sp. For e.g. some clones have a high uptake of Zn and Cu,

whereas others have a low uptake of Cd (Landberg & Greger 1994, Greger & Landberg 1999). This suggests, that the rhizosphere mechanisms regulate the release of metals from the soil colloids, and may be metal specific. Various clones of *Salix* spp. were cultivated in nutrient medium for three months. Subsequently these stem cuttings were transferred to a rhizobox system in which the root mass was connected to soil via a nylon net (25 µm) (figure 4). After 48-hours treatment, root exudates were collected from the rhizosphere by dipping the roots in redistilled water for 30 seconds. The solution was analysed for organic acids by ion chromatography, for Cu, Cd and Zn by AAS and for pH. The results showed that plant roots released organic acids viz. citric acid malic acid and succinic acid and changed the pH of the rhizosphere (Delhaize et al. 1993, Lee et al. 1977).

Arsenic is a toxic contaminant which is common in natural environments. Sorption/desorption of arsenate on/from different soil components may

Table 4. Possible processes fostering phytoremediation

Process fostering phytoremediation	Metal	Reference
Phytoextraction	Gold	Anderson et al. 1998
Metal accumulation	Gold, copper and zinc	Baker et al. 1988
Accumulation, phytovolatilization	Selenium	Bañuelos et al. 1991, 1992 a,b, 1993d, Bañuelos et al. 1997a,b, 1993c; Bañuelos and Meek 1990, Bañuelos et al. 1998, Brown and Shrift 1982, Lauchli 1993 ; Terry et al. 1992, Wu et al. 1988
Phytoremediation	Boron and selenium	Bañuelos et al. 1990, 1991, 1993a,b,
Accumulation and extraction	Boron	Bañuelos et al. 1995a,b, 2002
Root exudates and organic acids enhanced phytoremediatio	Aluminium	Barcelo and Poschenrieder 2002, Delhaize et al. 1993 Lee et al. 1977, Ma et al. 2001b, Mengoni et al. 2000
Accumulation	Air borne trace elements	Bargagli et al. 2003
Chelator enhanced uptake	Cadmium, Nickel and Chromium	Chen and Cutright 2001
Citric acid enhanced remediation	Toxic trace metals	Chen et al. 2003
Removal	Uranium	Dushenkov et al. 1997a
Phytoextraction	Zinc	Ebbs and Kochian 1998
Synthetic chelate enhanced phytoextraction	Lead	Huang 1997, Huang and Cunningham 1996 Huang et al. 1997
Phytoextraction	Toxic trace elements	Kumar et al. 1995
Phytoremediation	Uranium	Huang et al. 1998
Uptake	Radionuclide	Gouthu et al. 1997, Landeen and Mitchell 1986, Koranda and Robinson 1978
Bioavailability, root uptake, phytoremediation	¹³⁷ Cs	Lasat et al. 1997, 1998b, Fesenko et al. 1997b
Phytoextraction	Toxic metals	Lasat 2002
Hyper accumulation	Arsenic	Ma et al. 2001a
Hyper accumulator	Copper	Malaisse et al. 1978
Phytoextraction	Toxic trace elements	McGrath and Zhao 2003
Uptake, accumulation and translocation	Arsenate	Hartley-Whitaker et al. 2001, Macnair et al. 1992, Meharg and Macnair 1991, 1992
Uptake	Chromium(III) (VI)	Mishra et al. 1997
Availability and Uptake	Technetium99	Echevarria et al. 1977, Murphy and Johnson 1993
Hyper accumulation, tolerance	Cadmium	Nedelkoska and Doran 2000a, b
Bioavailability and bioaccumulation	Trace metals	Farago et al. 1977, Martin et al 1996, Prasad 1995
Accumulation	Nickel and zinc	Reeves 1988
Hyper accumulation	Lead and zinc	Reeves and Brooks 1983
Hyper accumulation	Nickel	Boyd and Martens 1994, Boyd et al. 1994, Brooks 1981, Dudley 1986a, b: Gabbrielli et al. 1990, 1991, Homer et al. 1991a,b, Jaffre et al. 1976, 1979, Kramer et al. 1997, Reeves et al. 1981, 1983, 1996, Robinson et al. 1997: Shetty et al. 1995.
Hyper accumulation	Zinc and cadmium	Brown et al. 1995a,b
Tolerance	Heavy metals	Arazi et al. 1999, Baker 1987Clemens 2001, El Aziz et al. 1991 Macnair 1993, Mengoni et al. 2000
Hyper accumulation	Copper and cobalt	Baker et al. 1983, Malaisse et al. 1978, 1979
Hyper accumulation	Zinc	Bashmakov et al. 2002

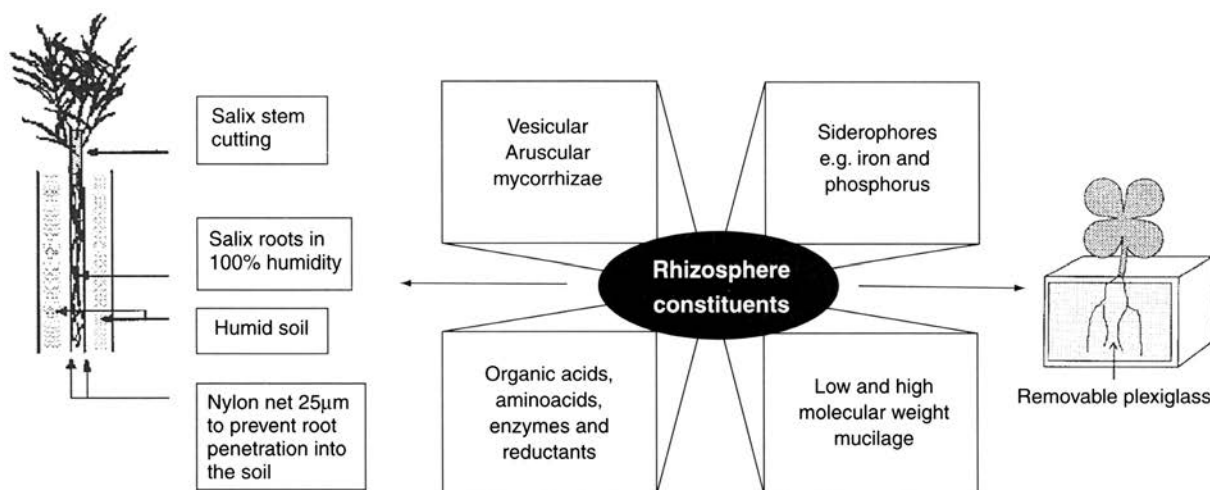


Figure 4 Rhizosphere constituents that would play a significant role in phytoremediation are amenable for collection and investigation using root box microcosm (redrawn based on Greger 2003)

have a dominant role in regulating arsenic mobility in soil-water-plant systems (Smith et al. 1998). To have information on the factors which may influence the mobility and potential toxicity of arsenic in soil environments, sorption/desorption of arsenate in the presence of inorganic phosphate, sulfate and low molecular mass organic ligands (LMMOLs) malate, oxalate, citrate and succinate has been investigated.

Organic ligands inhibited arsenate adsorption mainly at ligand/arsenate molar ratio > 1 , when added before arsenate and more in acidic than in neutral or alkaline environments. The efficiency of LMMOLs in preventing arsenate sorption on the Andisol at pH 4.0 was as follows: citrate $>$ malate $>$ oxalate $>$ succinate. The efficiency of organic ligands in preventing arsenate adsorption was also influenced by the nature and surface properties of variable charge minerals and soils (Pigna et al. 2003).

A novel approach by using synthetic cross linked polyacrylates and hydrogels has been examined on the metal-contaminated soils. A compound i.e. Stockosorb K400 when applied to hydrocultures of Scots Pine, *Pinus sylvestris*, which contained $1 \mu\text{M}$ of Pb, two useful functions have been observed: (i) the hydrogel increased the nutrient efficiency of the plants and, (ii) the detrimental effect of the heavy metal was completely remediated. Determination of the heavy metal content of the roots revealed that the uptake of the lead was greatly inhibited by the hydrogel. Analysis of the fine roots of 3- year-old spruce which had been grown for one vegetation

period in lead-contaminated soil with and without hydrogel amendments showed that, the amendment of the soil with the crosslinked polyacrylate and hydrogels did indeed prevent the uptake of the lead into the stele of the fine roots. These amendments inhibited the heavy metal uptake by the plant root. The presence of hydrogels (Stockosorb K 400) in the culture solution enhanced the root length and decreased lead concentration in the roots of *Pinus sylvestris*. (Hüttermann et al. 2004, in Prasad 2004) (figure 5). Lead can be made greatly more bioavailable by the addition of EDTA to soils lead, chromium and uranium can be removed by binding to soils and root mass via rhizofiltration (Dushenkov et al. 1995).

Bioconcentration of heavy metals by plants is well established and today a large number of higher plants that hyperaccumulate heavy metals are known (Brooks 1998 Dahmani-Muller et al. 2001). Plants exhibit strategies when exposed to elevated concentrations of heavy metals and they are variously classified as (i) indicators: plants in which uptake and translocation reflect soil metal concentration and show toxic symptoms (ii) excluders: plants that restrict the uptake of toxic metals into shoot over a wide range of background concentrations (iii) accumulators: plants in which uptake and translocation reflect background metal concentration without showing toxic symptoms and (iv) hyper-accumulators: plants in which metal concentration is up to 1% in dry matter (table 2). Thus, plants growing

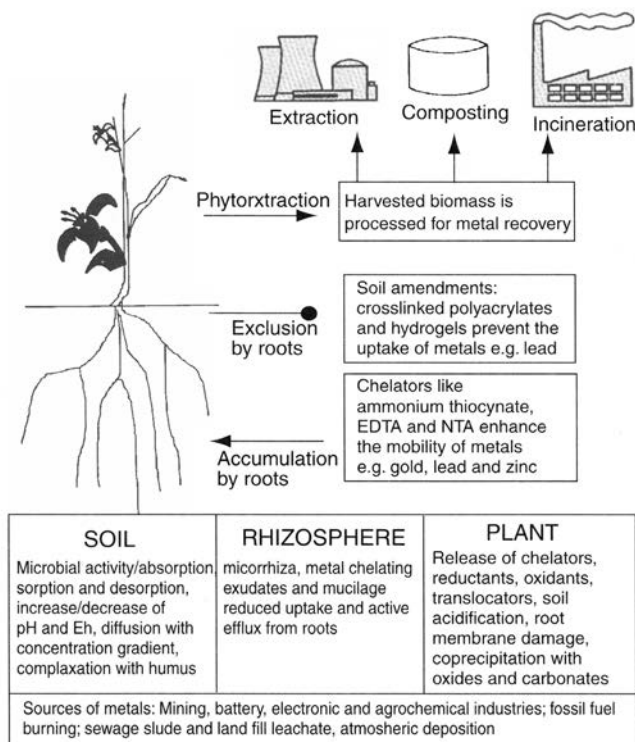


Figure 5. Phytoremediation for sustainable development. It consists of i) planting a species that tends to accumulate and store, transpire or degrade the target contaminant, ii) letting the crop grow and iii) harvesting it. Cofiring of phytomass or incineration and composting are concentration methods. Liquid extraction is another separation method. Variety of chelators have been tested to enhance phytoextraction. To achieve satisfactory results, the soil, rhizosphere and plant continuum processes need in-depth investigation

in metal contaminated and polluted terrestrial ecosystems take up toxic heavy metals and complex them in their parts, thus, help in environmental decontamination. Phytoextraction has been successfully applied for selenium (Bañuelos et al. 1999).

Phytomining uses hyperaccumulating plants to remove precious metals where conventional mining would not be economically feasible. The sulfur oxidizing bacteria have long been used in "biomining" to a significant extent for mining of copper, lead and nickel etc.

Several commercial agencies are pursuing research on phytomining. The pioneering phytoextraction lab of late Prof. Robert Brooks at Massey University in New Zealand discovered that *Brassica juncea* and other hyperaccumulating plant species accumulate as much as 20 ppm of gold in greenhouse pot experiments when plants were supplemented with the metal solubilizing

chelating agent such as ammonium thiocyanate. However, disposal of toxic cyanide, the leaching reagent of choice for gold and silver extraction in turn poses serious environmental problem. However, vascular plants (*Salix viminalis* x *schwerinii* & *Sorghum bicolor*) possessing appropriate enzymes are used as a part of phytoremediation to detoxify cyanide by converting it to the amino acid asparagine (Korte et al. 2000).

To date there are several commercial ventures gaining economic benefits from phytoextraction, not only by recovering extracted metals from plant biomass, but also in using the biomass for energy generation (Glass 1999, 2000). Plants that hyperaccumulate heavy metals, metalloids and radionuclides could potentially recontaminate the environment if not treated or processed appropriately. Any treatment or processing should result in a minimal volume of contaminant material requiring any further treatment. Two treatment methods, cofiring with coal/incineration and composting are concentration methods. The third, liquid extraction is a separation method. A subsample of harvested phytomass is burnt with coal under typical electrical power plant combustion conditions. Another fraction of phytomass was composted in containers and the third portion was subjected liquid extraction using chelation agents. Hyperaccumulators of metals contain leachable metals that would recontaminate the environment thereby requiring some type of post phytoremediation treatment. Cofiring with coal reduced the lead-contaminated mass by about 90% wt by concentrating the lead into small fly ash particles. Containers composting reduced the lead-contaminated material by about 26% wt, while extraction using chelating agents removed over 90% wt. of the lead in two sequential batch extractions. Hyperaccumulators have often been isolated from nature in areas of high contamination or high metal concentration. Among the earliest identified terrestrial hyperaccumulating species are members of the genus *Thlaspi* are known to accumulate Zn, Cd and Pb, and the *Alyssum* genus, which are Ni accumulators. For additional information reference be made to Prasad and Freitas 2003. Lemon-scented geranium (*Pelargonium* sp. 'Frensham') was reported to accumulate large amounts of Cd, Pb, Ni and Cu from soil in greenhouse experiments (Prasad 2004 in Prasad 2004a).

The commercial potential of hyperaccumulators derives from their rates of accumulation of metals, combined with their growth (i.e., biomass accumulation) rates: multiplying the metal accumulation rate (gm metal per kilogram of plant tissue) by the growth rate (kilograms of tissue per hectare per year), yields a value for metal removal from soil (gm or kg of metal per hectare per year). It has been estimated that this rate must be at least several hundreds and perhaps as much as 1,000 kg/ha/yr, to be commercially viable, and even then, remediation may take many years depending on initial metal concentrations and the depth of the contamination (Glass 1999, McGrath & Zhao 2003).

There are numerous reports in the literature of natural or selected species being able to accumulate larger concentrations of metal, in the range of 2-4% of dry weight. It has been estimated that accumulations of 1-2% might yield metal removal rates of 200-1,000 kg/ha/yr, depending on the growth rate of the plants, and thus be sufficient for commercial application.

MacGrath and Zhao (2003) discussed the potential of phytoextraction, since fast-growing hyper-accumulators that can produce a large biomass are known. However, higher growth and enhanced biomass production by most naturally occurring metal hyperaccumulators and the problem of limited solubility of metals in soils (i.e., the high affinity of metal ions for soil particles) are being addressed and there have been some significant developments in this area. The selection or creation of new plant varieties, including the use of genetic engineering tools have promising scope to enhance biomass production. Such efforts are also directed towards the use of artificially improved or selected varieties to enhance uptake of the metal contaminant (Cunningham & Ow 1996, Raskin 1996). The problem of limited solubility is being investigated through the use of chelators or other surfactants in the soil, or even combination of phytoremediation with other *in situ* techniques like electro-osmosis, to stimulate metal mobility in the soil.

Another strategy to enhance metal uptake by plants, being pursued by a number of research groups, is the use of symbiotic root-dwelling fungi. These fungi, called mycorrhizae, naturally live on the roots of many types of plants, and benefit their plant host by improving the plant's ability to take up water and certain types of nutrients, by

effectively increasing the surface area of the root available for such uptake (Wenzel et al. 1999 in Prasad & Hagemeyer 1999, Wenzel et al. 2004 in Prasad 2004)

In order for plant roots to take up metals from soil, the soil-bound metal must first be solubilized (i.e., freed from the organic or inorganic soil components to which they are bound). This could be accomplished by secretion of metal-chelating molecules or metal reductants by the plant roots (Raskin & Ensley 2000). Once solubilized, the metal ions may enter roots via extracellular or intracellular pathways, with the intracellular pathways generally requiring the presence of an ion channel or a metal transport protein in the plasma membrane of the root cells. It has been speculated that some of these channels may be nonspecific, so that they can be used by different metals (Raskin & Ensley 2000). Once metal ions reach the root, they can either be stored in vacuoles in the root, often in chelated form, or they can be transported to the shoots. It is reported that some hyperaccumulating species accumulate metals both in roots or shoots or both in root and shoot (Prasad 1998).

Phytoremediation has been applied to treat soil and water contaminated with arsenic, an element used in pesticides and wood treatment. Very well characterized crop species such as Indian mustard and sunflowers are currently in use for the phytoremediation technology. Researchers from all over the world have now shown that plants can accumulate lead, mercury, arsenic, chromium, uranium, cesium, strontium, gold, zinc, selenium, manganese, calcium, iron, magnesium and other metals from soils into harvestable plant parts. When coupled with techniques involving soil amendments and hyperaccumulation inducing agents, then plants can take up heavy metals more than 1% (up to 3% in some cases) of their dry weight.

A variety of different metals can be taken up by natural hyperaccumulators. The most readily bioavailable metals include Cd, Ni, Zn, As, Se and Cu, while more moderately bioavailable metals are Co, Mn and Fe, Pb, Cr and U are generally not considered very bioavailable, but all are capable of phytoremediation to some extent: certain plants can take up Pb, but availability is enhanced by the use of chelators.

Hyperaccumulators are capable of uptake and translocation of metal contaminants in the soil by plant roots into the above ground portions of the plants. After the plants have been allowed to grow

for several weeks or months, they are harvested and either incinerated or composted to recycle the metals. This procedure may be repeated as necessary to bring soil contaminant levels down to allowable limits. If plants are incinerated, the ash must be disposed of in a hazardous waste landfill.

Using *Brassica juncea* (Indian mustard), lead-contaminated shallow soils in Trenton, New Jersey have been remediated (Blaylock et al. 1997). Approximately 50% of the lead was removed from the surface soil (~700 mg/kg) in one year for phytoextraction to be effective, one needs vigorously growing plants (> 3 tons dry matter/acre/yr), an easily harvestable above-ground portion, and a plant that accumulates large amounts of metals (~1000 mg/kg) in above ground biomass. To achieve cleanup within three to five years, the plant must accumulate about ten times the level in soil (for example, if the level in soil is 500 mg/kg, then the concentration in the plant must be almost 5000 mg/kg to cleanup the soil in a few years). Some sites have metals that are bioavailable while others do not. Generally, cadmium, nickel, zinc, arsenic, and copper are relatively bioavailable while lead, chromium, and uranium are not taken up and translocated to the harvestable biomass. Plants which accumulate nickel, cobalt, copper, manganese, lead, zinc, and selenium have been reported in the literature (Brooks 1988). Zinc and boron are phytotoxic to some plants at levels above 200 mg/kg in soil. Addition of EDTA (0.5 to 10 kg EDTA/kg soil) has greatly enhanced the bioavailability of lead, but the enhancement must be weighed against the increased probability of lead migration to groundwater (Barona et al. 2001, Chen & Cutright 2001, Hong et al. 1999, Huang et al. 1997, Jiang et al. 2003, Vassil et al. 1998, Wu et al. 2003). Mathematical modeling of water movement and metals transport may be required to further understand the fate of lead under these conditions (Fiorenza et al. 2001).

Advantages

The plant biomass containing the extracted contaminant can be a resource. For example, biomass that contains selenium (Se), an essential nutrient, has been transported to areas that are deficient in Se and used for animal feed (Bañuelos et al. 1997a). In green house experiments, gold was harvested from plants (Anderson et al. 1998).

Limitations: (i) Metal hyperaccumulators are generally slow-growing with a small biomass and

shallow root systems (ii) Plant biomass must be harvested and removed, followed by metal reclamation or proper disposal of the biomass. Hyperaccumulators may accumulate significant metal concentrations e.g., *Thlaspi rotundifolium* grown in a lead-zinc mine area contained 8,200 g/g Pb (0.82%) and 17,300 g/g zinc (Zn) (1.73%), and *Armeria maritima* var. *halleri* contained 1,300 g/g Pb, dry weight basis (Baker et al. 1995 (iii) Metals may have a phytotoxic effect (Kumar et al. 1995) (iv) Phytoextraction studies conducted using hydroponically-grown plants, with the contaminant added in solution, may not reflect actual conditions of soil. Phytoextraction coefficients measured under field conditions are likely to be less than those determined in the laboratory (Kumar et al. 1995).

Phytostabilization

Establishment of rooted vegetation on mine tailings prevents windblown dust, an important pathway for human exposure at hazardous waste sites. Hydraulic control is possible, in some cases, due to the large volume of water that is transpired through plants which prevents migration of leachate towards groundwater or receiving waters. Metals do not ultimately degrade, so capturing them *in situ* is sometimes the best alternative at sites with low contamination levels (below risk thresholds) or vast contaminated areas where a large-scale removal action or other *in situ* remediation is not feasible. Vigorously growing plants are necessary to exert hydraulic control and immobilization at the site; plants cannot die or be removed during the phytostabilization design period. Low-level radionuclide contaminants can also be held in place by phytostabilization, and this alternative can result in significant risk reduction if their half-lives are not too long. Soil amendments such as phosphate, lime, and organic matter are sometimes needed to immobilize toxic metals such as lead, cadmium, zinc, and arsenic. Cadmium is readily translocated to leaves in many plants, which represents a risk to the food chain (Barman & Lal 1994, Brown et al. 1996, Chaney 1980, McLaughlin & Singh 1999) and this pathway may be the limiting consideration in applying phytostabilization at some metal contaminated sites. It involves the use of plants especially roots and/or plant exudates to stabilize, demobilize and bind the contaminants in the soil matrix, thereby reducing their bioavailability. This approach is

suitable for metal contaminated soils. Several plant species including common agricultural and horticultural crops have shown potential in phytostabilization. Field trails for *in situ* decontamination of heavy metal polluted soils using metal accumulating crops was proved to be advantageous, particularly with fibre yielding crops (Felix 1997, Prasad et al. 2003).

Phytostabilization relies on plants, or compounds they secrete, to stabilize low levels of contaminant that are present in soils (e.g., by absorption or precipitation), to prevent them from mobilizing or leaching in a manner that would endanger public health (Cunningham et al. 1995). Possible mechanisms might include sequestering the contaminant in or on cell wall lignin ("lignification"), absorption of contaminants to soil humus, via plant or microbial enzymes ("humification"), or other mechanisms whereby the contaminant is sequestered in the soil, e.g., by binding to organic matter. Phytostabilization is primarily applicable to metal contamination, and might best be used near the end of remediation by traditional means, or after site closure, for those sites where it is acceptable under regulatory guidelines to leave a non-bioavailable portion of the contaminant remaining in the soil (e.g. a risk-based regulatory approach). The term also refers to the use of a vegetative cover, e.g., at a landfill, to prevent contaminants from leaching into groundwater or surface waters (Glass 1999).

Phytostabilization reduces the mobility of the contaminant and prevents migration to the ground water or air, and it reduces bioavailability for entry into the food chain. (Barman & Lal 1994, Bañuelos & Mayland 2000, Bunzl et al. 2001, Chandra Sekhar et al. 2001, Dudka & Miller 1999, Dudka et al. 1996). This technique can be used to reestablish a vegetative cover at sites where natural vegetation is lacking due to high metal concentrations in surface soils or physical disturbances to surficial materials. Metal-tolerant species can be used to restore vegetation to the sites, thereby decreasing the potential migration of contamination through wind erosion, transport of exposed surface soils, and leaching of soil contamination to ground water (Kock & Mitchell 1957).

Phytostabilization occurs through root zone microbiology and chemistry, and/or alteration of the soil environment or contaminant chemistry. Soil pH may be changed by plant root exudates or through the production of CO₂. Phytostabilization can change

metal solubility and mobility or impact the dissociation of organic compounds. The plant affected soil environment can convert metals from a soluble to an insoluble oxidation state. Phytostabilization can occur through sorption, precipitation, complexation, or metal valence reduction. Plants can also be used to reduce the erosion of metal-contaminated soil. The term phytolignification has been used to refer to a form of phytostabilization in which organic compounds are incorporated into plant lignin (Cunningham et al. 1995). Compounds can also be incorporated into humic material in soils in a process likely related to phytostabilization in its use of plant material (Bollag & Myers 1992, Conte & Piccolo 2003).

Cattle provide remarkable Reclamation of mine tailings through the Development of Rhizosphere
In Globe Arizona, USA, copper mines rehabilitation and mine reclamation programme [Arizona Ranch, Resource Management and Mine Reclamation and Tucson (ASARCO Inc. Copper Operations)] is primarily based on cattle. Due to stringent environmental legislation the mine operators are under the obligation of restoration of metal polluted and contaminated sites. In order to curb the metal pollution several of the metal polluting industries and agencies are looking for low cost and effective measures rather than adopting conventional chemical methods which are cost prohibitive. To achieve mine reclamation, herds of cattle are impounded with electric fence on the sites of mine tailings, metal contaminated soils by providing fodder and water for varying durations. Cattle stabilize the soil by their hoofs and enrich nutrient status through dung and urination. This process is repeated at regular intervals in cycles often

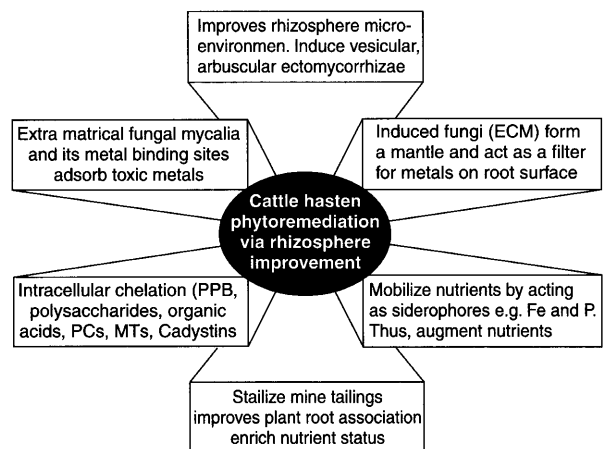


Figure 6 Cattle hasten bioremediation of mine tailings by improving the rhizosphere environment.

spreading for years. This exercise provided convincing evidence for cattle accelerated rhizosphere development and enriched soil microorganisms and nutrients (figure 6) (Dagget 1997).

Advantages: (i) It circumvents the removal of soil, (ii) It has a lower cost and is less disruptive than other more-vigorous soil remedial technologies, (iii) Revegetation enhances ecosystem restoration.

Limitations: (i) The contaminants remain in place. (ii) The vegetation and soil may require long-term maintenance to prevent re-release of the contaminants and future leaching, (iii) Vegetation may require extensive fertilization or soil modification using amendments (Zorpas et al. 1999), (iv) Plant uptake of metals and translocation to the aboveground portion must be avoided, (v) The root zone, root exudates, contaminants, and soil amendments must be monitored to prevent an increase in metal solubility and leaching, (vi) Phytostabilization might be considered to only be an interim measure, (vii) Contaminant stabilization might be due primarily to the effects of soil amendments, with plants only contributing to stabilization by decreasing the amount of water moving through the soil and by physically stabilizing the soil against erosion.

Plants that produce root exudates are shown to stimulate growth of beneficial microorganisms or stimulate cometabolism will be of more benefit than plants without such useful exudates (Barcelo et al. 2003, Barcelo and Poschenrieder 2002, Yoshitomi & Shann 2001). The type, amount, and effectiveness of exudates and enzymes produced by 'plantroots' will vary between species and even within species or varieties of one species.

Phytovolatilization

This method can be used only for those contaminants that are highly volatile. Mercury or selenium, once taken up by the plant roots, can be converted into non-toxic forms and volatilized into the atmosphere from the roots, shoots, or leaves. For examples, Se can be taken up by plants of the *Brassica* genus and other wetland plants, and converted (for example, by methylation to the volatile dimethyl selenium) into non-toxic forms which are volatilized by the plants: field testing has shown this to be a potentially effective method. A similar mechanism can be exploited for Hg, although there are no naturally occurring plants that can accomplish this: the goal here is to

engineer bacterial genes for mercury reduction into plants, and here too laboratory experiments are highly encouraging (Bizily et al. 1999, Heaton et al. 1998).

Phytovolatilization occurs as growing trees and other plants take up water and the contaminants. Some of these contaminants can pass through the plants to the leaves and volatilize into the atmosphere at comparatively low concentrations. Mercury has been shown to move through a plant and into the air in a plant that was genetically altered to allow it to do so. The thought behind this media switching is that elemental Hg in the air poses less risk than other Hg forms in the soil.

Advantages: (i) Contaminants could be transformed to less-toxic forms, such as elemental mercury and dimethyl selenite gas, (ii) Contaminants or metabolites released to the atmosphere might be subject to more effective or rapid natural degradation processes such as photodegradation.

Limitations: The contaminant or a hazardous metabolite might accumulate in vegetation and be passed on in later products such as fruit or lumber. Low levels of metabolites have been found in plant tissue.

Rhizofiltration

Rhizofiltration has been pioneered by Dr Ilya Raskin and the group at Rutgers University, NJ (USA) (Dushenkov et al. 1995, Kumar et al. 1995). It is effective in cases where wetlands can be created and all of the contaminated water allowed to flow through inclose contact with roots. Contaminants should be those that sorb strongly to roots such as, lead, chromium(III), uranium, and arsenic (V).

Rhizofiltration refers to the use of plant roots to sorb, concentrate, and precipitate metal contaminants from surface or groundwater. Roots of plants are capable of sorbing large quantities of lead and chromium from soil, water or from water that is passed through the root zone of densely growing vegetation. Shallow lagoons have been engineered as wetlands and maintained as facultative microbial systems with low dissolved oxygen in the sediment. Groundwater or wastewater is pumped through the system for the removal of contaminants by rhizofiltration. Wetlands have been used with great success in treating metals for many years (McGrath et al. 2001, Wenzel et al. 2003, Fitz & Wenzel 2002). Long-term utilization of wetland plants and sulfate-reducing conditions result in an increase in pH and a decrease in toxic metals concentrations for treatment of acid mine drainage. Root systems and sediments

in wetlands are facultative (aerobic and anaerobic zones) which facilitates sorption and precipitation of toxic metals. (Dushenkov et al. 1995, Fitz & Wenzel 2002). Experimental evidence showing nonlinear kinetics of disappearance of metals from solution suggests that several different mechanisms operate simultaneously (Kumar et al. 1995). Surface absorption by the roots, the fastest and often the most prevalent mechanism, most likely depends on physico-chemical processes (e.g., ion exchange, chelation) and can even take place on dead roots.

Rhizofiltration is the use of plant roots to accumulate metals from water. Hydroponically cultivated plants rapidly remove heavy metals from water and concentrate them in the roots and shoots. Harvested plants containing heavy metals can be disposed of or treated to recycle the metal. Today scientists have identified plants demonstrating high biomass production and metal removal capacity for a wide variety of metals. Rhizofiltration has many of the benefits of other phytoextraction techniques, including low cost and minimal environmental disruption. A continuous flow system circulates the contaminated water through specially designed plant containment units. Periodically, older plants are harvested and replaced.

In its reliance on surface absorption as the primary mechanism for removing metals from waste streams, rhizofiltration is related to the process known as biosorption, in which microbial, fungal or other biomass, living or dead, is used to absorb large quantities of materials such as heavy metals. In addition to surface absorption, other, slower mechanisms underlying rhizofiltration may also occur: these might include biological processes (intracellular uptake, deposition in vacuoles, and translocation to the shoot), or precipitation of the metal from solution by plant exudates (the slowest mechanism).

There are several naturally occurring plant species with a natural propensity to accumulate high concentrations of metal. The desired species are those where root surface absorption is the primary, or fastest mechanism: in those plants where a substantial amount of shoot translocation occurs, more of the plant biomass becomes contaminated with the metal or radionuclide and must therefore be disposed as hazardous (or radioactive) waste, thus negatively affecting economics and efficiency. Rhizofiltration is believed to be effective (and perhaps most economically attractive) for dilute

concentrations of contaminants in large volumes of water, and this feature may make it especially attractive for radionuclide contamination (Dushenkov et al. 1995, Kumar et al. 1995).

Rhizofiltration can be practiced *in situ* (e.g., on surface waters), but more likely would be operated in an aboveground artificial flow-through reactor, akin to hydroponics, and could thus be applicable to groundwater or industrial waste streams. At the conclusion of the process, plant roots and any other tissue containing metals can be harvested for disposal or metal recovery: any uncontaminated shoots can, in some cases, be used to regenerate new roots.

Advantages: (i) Either terrestrial or aquatic plants can be used. Although terrestrial plants require support, such as a floating platform, they generally remove more contaminants than aquatic plants, (ii) This system can be either *in situ* (floating rafts on ponds) or *ex situ* (an engineered tank system), (iii) An *ex situ* system can be placed anywhere because the treatment does not have to be at the original location of contamination.

Limitations: (i) The pH of the influent solution may have to be continually adjusted to obtain optimum metals uptake, (ii) The chemical speciation and interaction of all species in the influent have to be understood and accounted for, (iii) A well-engineered system is required to control influent concentration and flow rate, (iv) The plants (especially terrestrial plants) may have to be grown in a greenhouse or nursery and then placed in the rhizofiltration system, (v) Periodic harvesting and plant disposal are required, (vi) Metal immobilization and uptake results from laboratory and greenhouse studies might not be achievable in the field.

Biodiversity prospecting for Phytoremediation and Metal Exploration

"Biodiversity prospecting" offers several opportunities of which the most important is to save as much as possible of the world's immense variety of ecosystems (Myers 1990). Biodiversity prospecting would lead to the discovery of wild plants that would not only clean metal polluted environments but also would help in mineral exploration (Aery & Tiagi 1986, Bargagli et al. 2003, Brooks et al. 1977a,b, Brooks 1972, Cannon 1959, 1960). Published literature indicate that wide variety of plants have been tested in field and laboratory (Prasad 2001). Remediation programmes relying on metal tolerant plants would be successful (Glass 1999, 2000).

The most successful monitoring methods for metals in the environment are based on bacterial heavy metal biosensors viz., a) gene based biosensors and b) protein based biosensors (Prasad 2001). Mosses, liverworts and ferns are also capable of growing on metal-enriched substrates. These plants possess anatomical and physiological characteristics enabling them to occupy unique ecological niches in natural metalliferous and man made environments. For example, groups of specialized bryophytes are found on Cu enriched substrates; so-called 'copper mosses' and come from widely separated taxonomic groups. Other bryophytes are associated with lead and zinc enriched substrates. However, the information about bryophytes growing on serpentine soils is rather scanty (Prasad 2001). Pteridophytes (ferns) are associated with serpentine substrates in various parts of the world. *Pteris vittata* (Brake fern), a fast growing plant is reported to tolerate soils contaminated with arsenic as much as 1500 ppm and its fronds concentrate the toxic metal to 22,630 ppm in 6 weeks (Ma et al. 2001a, Tu et al. 2003). Among angiosperms, about 400 metal hyperaccumulators have been identified which would serve as a reservoir for biotechnological application (Brooks 1998).

Brassicaceae had the highest number taxa i.e. 11 genera and 87 species that are established for hyper accumulation of metals (In Brassicaceae Ni hyper accumulation is reported in 7 genera and 72 species; Zn in 3 genera and 20 species) (Palmer et al. 2001, Prasad & Freitas 2003).

Considerable progress had been achieved recently in unraveling the genetic secrets of plants that hyper accumulate metals including genetic engineering for metal tolerance (Gisbert et al. 2003, Heaton 1998, Karenlampi et al. 2000, Macnair 1993, Macnair et al. 1992, Meharg & Macnair 1992, Mengoni et al. 2000, Peña & Séguin 2001, Pollard et al. 2002, Pollard & Baker 1996, Raskin 1996, Schat & Bookum 1992a). Genes responsible for metal hyperaccumulation in plant tissues have been identified and cloned (Moffat 1995, 1999). These findings are expected to identify new non-conventional crops, metallo crops that can decontaminate metals in the environment (Ebbs & Kochian 1998, Madejon et al. 2003, Raskin 1996). The fundamental aspects of microbe/plant stress responses to different doses of metals coupled with breakthrough research innovations in biotechnology would successfully provide answers as to how to apply biodiversity for advancing phytoremediation technology.

Many of the hyperaccumulators inhabit in serpentine soils as well as in a wide range of environmental conditions. In addition to serpentine-phytes, a number of wild brassicaceae are the best suitable species being hyperaccumulators for phytoremediation. Serpentine habitats and species are threatened worldwide. Mainly due to habitat loss, several serpentine endemic species have become extinct or they are highly threatened. These habitats, often regional hotspots of biodiversity should be preserved and actions to preserve these unique spots should be promptly implemented (Prasad & Freitas 1999, 2003).

Phytoremediation Research for Sustainable Development

Evaluation of promising agricultural crops e.g. kenaf and brassica based phytoremediation systems might be advantageous to Indian scenario for phytoextraction (Bañuelos et al. 1999, 2002, Ebbs & Kochian 1998, Linger et al. 2002), development of tree crop based (e.g. *Eucalyptus*, *Prosopis*, *Acacia* etc.) phytoremediation technology would be desirable (Bhati & Singh 2003, Greger & Landberg 1999, Latimer et al. 1996, Eklund 1995, El-Hasan 2003, Hagemeyer et al. 1994, Landeen & Mitchell 1986, Pulford & Watson 2003). A community of tree, legume and grass based phytoremediation systems would be advantageous (Barlow et al. 2000, Piechalak et al. 2002, Sahi et al. 2002) (figure 7).

Identification of low cost, effective and environmentally friendly chelators for enhanced accumulation of Pb and other metals in Indian mustard and other promising crops (Blaylock et al. 1997), implications of use of chelators on ground water, (Barona et al. 2001) and development of

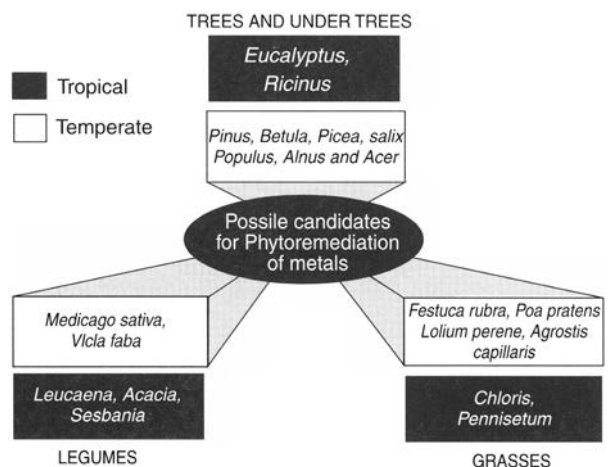


Figure 7 Possible candidates for phyto remediation of metals in temperate and tropical ecosystems

refined techniques for separating heavy metals from the phytomass of hyperaccumulators (Lakshmi & Clements 2003) needs critical investigation.

Application of plant tissue culture and plant biotechnological tools such as somatic hybridization between compatible heavy metal hyperaccumulators, and development of transgenic plants (Brewer et al. 1997), development of promising hairy root cultures for rhizofiltration of radionuclides (Eapen et al. 2003, Gleba et al 1999), transgenic crops for promising and low cost phytoremediation (Gisbert et al. 2003, Pilon-Smits & Pilon 2000, 2002) (figure 8).

Identification of indigenous plants that hyperaccumulate heavy metals and eventually leading to phytomining (Botanical prospecting of ore deposits) (Brooks 1998, Brooks et al. 1998, Cannon 1960). Detailed studies on indigenous Brassicaceae for metal accumulation (Palmer et al. 2001), identification of indigenous sources of biomass for removal of specific metals from aqueous solutions (Prasad et al. 2001), identification of indigenous clays for sorption of metal contaminants (Churchmann et al. 1999), identification of plants capable of accumulating airborne elements would be worth attempting (Bargagli et al. 2003, El-Hasan et al. 2003, Keane et al. 2001). Analysis

and examination of indigenous grass species inoculated with mycorrhizal fungi for their ability to accumulate metals, metalloids and radio nuclides (Entry et al. 1999, Khan 2001, Khan et al. 2000, Dushenkov et al. 1995, 1997a,b, 1999, Hossner et al. 1998), examination of arbuscular mycorrhizal associations for remediation of metalliferous and coal fly ash (Bi et al. 2003, Veecken & Hamelers 2002 2002), plant community studies for tolerant to toxic trace metals for restoration of the degraded mines soils are warranted (Dahmani-Muller et al. 2000, Freitas et al. 2003, Wong 2003). In-depth analysis of indigenous ferns for arsenic hyper accumulation (Francesconi et al. 2002, Ma et al. 2001a, Zhang et al. 2002) would contribute to sustainable development of metalliferous soils.

Studies on the molecular mechanisms (molecular farming) of plant metal tolerance (Clemens 2001, Clemens et al. 2002), expression of microbial genes from promising organisms e.g. *merB* in *Arabidopsis thaliana* that confer resistance to organo mercurials (Bizily et al. 1999), and production of genetically transformed trees over expressing 3-glutamyl cysteine synthetase will have potential role in phytoremediation (Arisi et al. 2000), role of phytochelatins and metallothioneins (figure 9) (Briat & Lebrun 1999, Cobbett & Goldsbrough 2002, Prasad 2003a,b), plant ABC transporters (Guerinot 2000, Prasad 2002a, 2003a, Theodoulou 2000, Karenlampi 2000), role of citric (Chen et al. 2003), humic and fulvic acids would be the promising fields of investigation for phytoremediation of metal contaminated soils (Conte & Piccolo 2003).

Metal ions are complexed in the cytosol with glutathione (GSH) and the derived PCs are

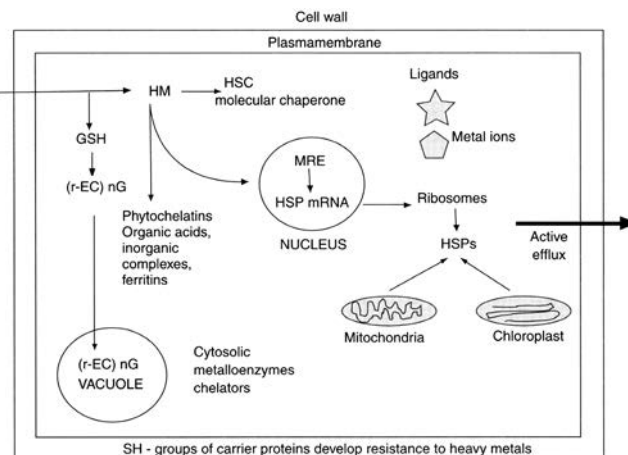
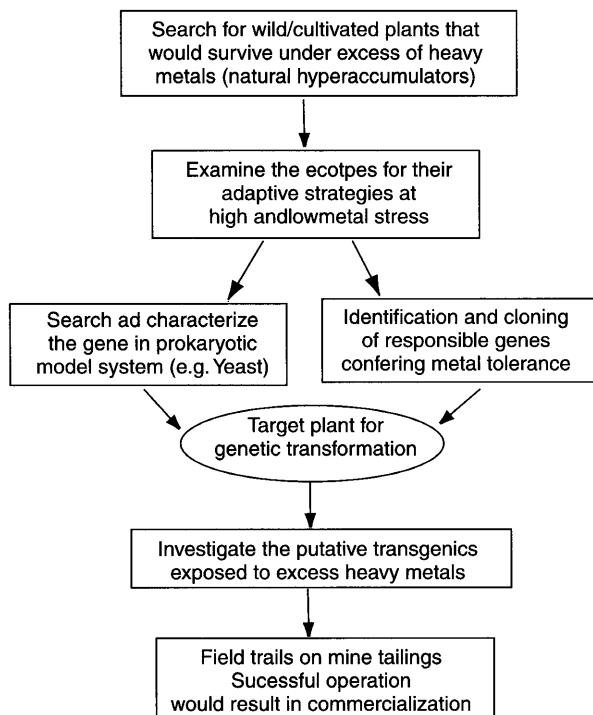


Figure 9 Cellular responses of heavy metal toxicity the physiological and molecular basis for the underlying mechanisms of phytoremediation

Figure 8 Feasible scheme for genetic improvement and transgenic production of plants for phytoextraction of metals

transported into vacuole. These functions have been implicated in heavy metal complexation and transport (Prasad 1999, Rauser 1995, Reddy & Prasad 1990). Another cellular response observed was that heavy metals interact with metal regulating elements (MRE) at the promoter region located in MT genes. For example in soybean such sequences have been found in the upstream region of heat-shock gene coded for 17.5 kDa HSP. Some heat-shock genes and heat shock proteins (HSPs) have been induced by Cd ions in soybean. These HSPs have been implicated as molecular chaperones (Neumann et al. 1994, Reddy & Prasad 1993, Prasad 1997, Prasad 1999).

Studies on the ligands exuded by plants in rhizosphere (Deiana et al. 2003, Prasad 1998), rhizosphere biotechnology of indigenous plants to facilitate metal accumulation (Barceló et al. 2003, Deiana et al. 2003, Fitz & Wenzel 2002, Jauert P et al. 2002, McGrath et al. 1997, 2001, Mishra et al. 1997, Tepfer et al. 1989, Wackett & Allen 1995, Walker et al. 2003, Wenzel et al. 2003 (figure 4), plants with concurrent ability for uptake of heavy metals and persistent organic pollutants from soil (Mattina 2003) through soil-rhizosphere-plant continuum (figure 5), accumulation of potentially toxic elements in edible plants and their transfer to human food chain (Barman & Lal 1994, Dudka & Miller 1999, Yusuf et al. 2003) need focused research.

Development of modeling for environmental phytoremediation (Blaylock et al. 1995, Brennan & Shelley 1999, Fesenko et al. 1997a, Fiorenza et al. 2001, Ouyang 2002), identification of plant and animal indicators that might serve a realistic tool are warranted (Beeby & Richmond 2002, Bañuelos & Mayland 2000).

Species selective analysis and localization of metals in plant cell environment by sophisticated analytical tools would be advantageous to localize metals in plant cells, however expensive they might be such as: atomic absorption spectrometry AAS, atomic emission spectrometry AES, atomic fluorescence spectrometry AFS, anodic-stripping voltammetry ASV, energy dispersive X-ray fluorescence EDXRF, electro thermal atomic absorption spectrometry ETAAS, extended X-ray absorption fine structure spectroscopy EXAFS, flame AAS, flame AES flame AFS, particle-induced gamma-ray emission PIGE, particle-induced X-ray emission PIXE (Ager et al. 2002, Montes-Bayon et al. 2002, Schaumlöffel et al. in Prasad 2004a).

Conclusions

In spite of certain limitations, there are several

advantages of phytoremediation technology. There have been apprehensions and often metal accumulators are considered as if the contaminants are only transferred and are potentially toxic when the plants accumulate these contaminants. In this regard, industrial crops not used for food production e.g. fibre crops would be the best alternative land use option. Industrial crops such as *Brassica juncea*, *Armoracia rusticana*, *Arabidopsis halleri*, *Gossypium hirsutum*, *Helianthus annuus*, *Eucalyptus*, *Amaranthus*, *Cannabis sativa* and *Linum usitatissimum* based phytoremediation systems would contribute to sustainable development.

Agricultural production in metal/radionuclide contaminated soil is of human health concern because of high levels of metallic and radionuclide residues in agricultural food products. Therefore, interest arises in developing more integrated and ecologically-based approaches. In this context, industrial crops not used for food production such as fibre crops are thought of as an alternative option. Two fibre crops viz, hemp (*Cannabis sativa* L.) and flax (*Linum usitatissimum* L.) were preferred for cultivation (Linger et al. 2002, Vandehove & Van Hees 2003). For both crops extensive knowledge is available on the use of the fibre for textiles and many other high value applications. Hemp has been investigated as a biofuel for energy production through thermal conversion and scores high on ecological and economic criteria. However, information on the radionuclide transfer to fibre crops is limited (Bañuelos et al. 1995a, 2002).

Higher soil potassium (K) levels decrease the Cs-transfer factor (TF), therefore, appropriate K-fertilisation will result in a decreased TF in most cases. Lower TF's were observed for flax than hemp, but the TFs to the usable plant parts both for hemp and flax, are low enough to allow for the production of a clean end product (fibre, seed oil, biofuel) even on heavily contaminated land (1480 kBq m⁻²). Flax is a more demanding crop compared to hemp in terms of soil properties (more fertile, fine textured soils) and agronomy. Depending on soil type, climate conditions and requirements for final product use, one or the other crop will be an appropriate alternative land use (Linger et al. 2002, Vandehove & Van Hees 2003). Thus, fibre and energy crops and industrial crops that are amenable to genetic manipulation via *in vitro* culture techniques would certainly play significant role for the success of phytoremediation for sustainable development.

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